ACCUMULATION OF CADMIUM IN THREE SUNFLOWER (HELIANTHUS ANNUUS L.) CULTIVARS

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Abstract

The effects of different concentrations of Cadmium on the growth of sunflower (Helianthus annuus L.) and its accumulation in roots, stem and leaves were investigated using inductively coupled plasma atomic emission spectrometry (ICP-AES). The concentrations of Cadmium chloride ranged from 10^{-5} M to 10^{-3} M. Seedlings of three sunflower cultivars viz., No. 665, RH118 and QFS14 exposed to 10^{-3} M Cd exhibited substantial growth reduction and all of them died 10 days after treatment application. Growth of roots and shoots was inhibited at concentrations of 10^{-5} M and 10^{-4} M Cd during the entire experiment (20 days). Cadmium accumulation in roots, stems and leaves increased significantly ($p<0.05$) with increasing Cd concentration. Cadmium was concentrated mainly in the roots, and variable amounts of Cd were also transported to stem and leaves. Among the three cultivars, RH118 produced more roots and higher biomass than No. 665 and had a greater ability to accumulate Cd when compared with QFS14.

Introduction

Cadmium (Cd) a trace heavy metal is considered a hazardous pollutant due to its high toxicity and solubility in water (Lockwood, 1976). At low concentration it has stimulatory effect on root growth of Allium sativum (Liu et al., 2003/2004), but at higher concentration it is toxic and directly or indirectly inhibits physiological processes such as respiration, photosynthesis, cell division, plant-water relationships, N metabolism and mineral nutrition, resulting in poor growth and low biomass (Barceló & Poschenrieder, 1990; Sanità & Gabbrielli, 1999; Liu et al., 2003/2004). Cadmium can be easily taken up by plants, transferred through food chains to impart adverse effects on human health (Nordberg, 2003; Wagner, 1994).

Emphasis has become more prevalent towards the problems of Cd pollution with the development of modern industry and agriculture. Anthropogenic activities, such as mining, industry, agriculture and waste disposal has increased since the beginning of the 20\textsuperscript{th} century (Alloway, 1995). Most conventional remediation approaches do not provide an acceptable solution to treating contamination caused by metal pollution. Alternatively, phytoremediation provides an attractive strategy, being low cost and environmentally sustainable (McGrath et al., 2002; Salt et al., 1998). The idea of using rare plants which hyperaccumulate metals to selectively remove and recycle excessive soil metals was introduced by Chaney (1983). Salt et al. (1996) also considered phytoremediation as an emerging technology using selected and engineered metal-accumulating plants for environmental clean-up. Metal-accumulating plants can accumulate unusually high concentrations of heavy metals in both roots and shoots from polluted soil and waters (Kumar et al., 1995; Dushenkov et al., 1995). A few terrestrial plant species e.g. Thlaspi sp., (Lombi et al., 2000) and Arabidopsis halleri (Bert et al., 2002) were reported to accumulate high concentrations of Cd, higher than the hyperaccumulation criterion, 0.01\% of shoot dry matter (Baker et al., 2000).

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Sunflower (Helianthus annuus), a fast-growing crop, has a reasonable tolerance to heavy metals. It has been used for rhizofiltration because it has a high root uptake of metals but shows low efficiency in their translocation from root to shoot (Saxena et al., 1999; Kamnev & Van der Lelie, 2000; Lin et al., 2003).

Few reports on cadmium accumulation by H. annuus are available (Madejón et al., 2003; Pena et al., 2006). The aim of this investigation was to study the effects of different Cd concentrations on growth of sunflower. In addition, the accumulation of Cd by sunflower roots, stem and leaves was measured using inductively coupled plasma atomic emission spectrometry (ICP-AES).

Materials and Methods

Plant cultivation and treatment application: Seeds of three sunflower (Helianthus annuus L.) cultivars No. 665, RH118 and QFS14 used in the present investigation were provided by Qinfeng Agricultural Science Company in Yangling, Shanxi province, P.R. China.

Healthy and equal-sized seeds were chosen from each cultivar, soaked in tap water for 12 h and germinated in Petri dishes in the dark at a constant temperature of 25°C for 48 h. Seeds were allowed to grow in vermiculite in a climate chamber with a day/night period 14/10 h, a day/night temperature and humidity regime of 20°C/25°C and 55/75% relative humidity (RH), respectively. After 7 days 20 sunflower seedlings were selected, fixed on a polystyrol-plate in a pot containing 2L of 1/2 modified Hoagland nutrient solution (Stephan & Prochazka, 1989) for 10 days, respectively. The Hoagland’s solution consisted of 5 mM Ca(NO₃)₂, 5 mM KNO₃, 1 mM KH₂PO₄, 50 µM H₃BO₃, 1 mM MgSO₄, 4.5 µM MnCl₂, 3.8 µM ZnSO₄, 0.3 µM CuSO₄, 0.1 µM (NH₄)₆Mo₇O₂₄ and 10 µM FeEDTA adjusted to pH 5.5. Seeds were nest treated with different concentrations of Cd ranging from 10⁻⁵ M to 10⁻³ M for 20 days. Cadmium was provided as Cadmium chloride. The Cd solutions were prepared in deionized water and were added to the full strength Hoagland nutrition solution. The full strength Hoagland solution without Cd was used for the control plants. The nutrient solutions were continuously aerated and changed regularly every 5 days until the seedlings were harvested.

Macroscopic observations: The lengths of roots and stems were measured after 5, 10, 15 and 20 d, respectively. Ten seedlings from each treatment were harvested based on uniformity in size and colour after 20 days of incubation. The seedlings were removed from solution and washed thoroughly with running tap water for 30 min. and then with deionized water to remove traces of nutrients and Cd ions from root surfaces. The samples were dried for 3 days at 45°C, for 1 day at 80°C, and for 12 h at 105°C.

Estimation of total Cd: All dried plant samples were prepared using a wet-digestion method (Piper, 1942). Tissue Cd concentrations were analyzed using inductively coupled plasma atomic emission spectrometry (ICP-AES, LEEUMAN LABS INC., New Hampshire, USA) as described by Duan (2003).

Statistical analysis: Data were expressed as mean±standard error (SE) and were analyzed through analysis of variance (ANOVA) using Sigma statistical software (Jandel Scientific Corporation, USA). Test of equality of averages using a t-test was applied equally. The statistical significance was set at p<0.05.
Table 1. Cadmium concentration (µg/g DW) and distributive changes in roots, stem and leaves of Helianthus annuus L.

<table>
<thead>
<tr>
<th>Cultivars</th>
<th>Treatment (M)</th>
<th>Total amount (%)</th>
<th>Root (%)</th>
<th>Stem (%)</th>
<th>Leaf (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Control</td>
<td>15.8 (100)</td>
<td>5.0 ± 0.73a</td>
<td>4.6 ± 0.61a</td>
<td>4.2 ± 0.66a</td>
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<tr>
<td>No. 665</td>
<td>10^{-5}</td>
<td>762.1 (100)</td>
<td>623.2 ± 5.7b</td>
<td>73.8 ± 2.42b</td>
<td>65.1 ± 2.73b</td>
</tr>
<tr>
<td></td>
<td>10^{-4}</td>
<td>2742.7 (100)</td>
<td>1878.4 ± 168.33c</td>
<td>605.1 ± 4.45c</td>
<td>259.2 ± 21.16c</td>
</tr>
<tr>
<td></td>
<td>10^{-3} No data</td>
<td>No data</td>
<td>No data</td>
<td>No data</td>
<td>No data</td>
</tr>
<tr>
<td>RH118</td>
<td>Control</td>
<td>3.8 (100)</td>
<td>2.3 ± 0.19a</td>
<td>0.4 ± 0.08a</td>
<td>1.1 ± 0.15a</td>
</tr>
<tr>
<td></td>
<td>10^{-5}</td>
<td>693.9 (100)</td>
<td>544.7 ± 13.12b</td>
<td>84.6 ± 1.49b</td>
<td>64.6 ± 1.61b</td>
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<tr>
<td></td>
<td>10^{-4}</td>
<td>2278.5 (100)</td>
<td>1315.4 ± 13.40c</td>
<td>617.6 ± 2.61c</td>
<td>345.5 ± 6.02c</td>
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<td>No data</td>
<td>No data</td>
<td>No data</td>
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<tr>
<td>QFS14</td>
<td>Control</td>
<td>0.8 (100)</td>
<td>0.5 ± 0.06a</td>
<td>0.1 ± 0.10a</td>
<td>0.1 ± 0.13a</td>
</tr>
<tr>
<td></td>
<td>10^{-5}</td>
<td>469.9 (100)</td>
<td>316.3 ± 2.10b</td>
<td>71.2 ± 0.65b</td>
<td>82.5 ± 2.81b</td>
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<tr>
<td></td>
<td>10^{-4}</td>
<td>2030.0 (100)</td>
<td>1150.1 ± 29.76c</td>
<td>511.1 ± 11.60c</td>
<td>368.8 ± 2.79c</td>
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<tr>
<td></td>
<td>10^{-3} No data</td>
<td>No data</td>
<td>No data</td>
<td>No data</td>
<td>No data</td>
</tr>
</tbody>
</table>

Values followed by same letters are not significantly different at (p>0.05). Means ± SE, n= 3. No data: all seedlings died at 10^{-3} M Cd. Values in parentheses are percent of control.

Results

Macroscopic symptoms: The effects of Cd on root growth of cultivars varied with Cd concentration (Fig. 1 A-C). Cadmium inhibited root growth during the entire experiment, when compared with control seedlings. Root growth in sunflower cultivar No. 665 was inhibited significantly, and root length decreased with increasing Cd concentration and duration of treatment (Fig. 1 A). Seedlings exposed to 10^{-3} M Cd solution exhibited substantial growth reduction, and root growth even stopped; they appeared thinner and were sparsely branched. The roots became yellow and were slightly decomposed after 2 days of treatment. After 5 days, the roots appeared yellow brown and some were broken. Root growth in the other two cultivars was approximately the same as that of No. 665 (Fig. 1B and C). Biomass of roots was in the order: QFS14 > RH118 > No. 665. The root system of QFS14 appeared qualitatively strongest.

The effects of Cd on shoot growth varied with different Cd concentrations (Fig. 2 A-C). Shoot length in the three cultivars progressively decreased with increasing Cd concentration and duration of treatment. The least shoot growth was in seedlings treated with 10^{-3} M Cd. Biomass of shoots was in the order: QFS14 > RH118 > No. 665. Control seedlings had 5 pairs of leaves when they were harvested. The plants of No. 665 treated with Cd did not grow as large, robust or healthy as the control during the experiment. With increasing Cd concentration, this phenomenon appeared progressively obvious. After 10 days the young laminas exposed to 10^{-5} M and 10^{-4} M Cd solutions appeared yellow. After 15 days chlorosis was evident in all young laminas. Toxic symptoms began in 10^{-3} M Cd 2 days after treatment and after 5 days chlorosis and necrosis appeared. The shoots wilted and leaf margins and leaf apices also wilted. When compared to No. 665, the toxic effects of Cd on laminas of the other two cultivars were less severe.
At 10^{-3} M Cd, the root and shoot length significantly decreased and the seedlings died after 10 days (Figs. 1 and 2). These phenomena indicated sensitivity of all root and shoot growth to higher concentrations of Cd (10^{-3} M).

**Accumulation of Cd:** The accumulation of Cd in the roots, stem and leaves of the three cultivars increased significantly (p<0.05) with increasing Cd concentration. Cd ions accumulated primarily in roots (Table 1). The Cd concentration in roots of No. 665 was higher than that in RH118 and QFS14 at 10^{-4} M Cd. Cadmium was also transported to
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The Cd shoot/root ratios increased with increasing Cd concentration. Levels of Cd in the three cultivars treated with $10^{-5}$ M and $10^{-4}$ M Cd were in the order: roots > stem > leaves, with the exception of roots > leaves > stem at $10^{-5}$ M Cd in QFS14. In contrast, No. 665 showed the best capacity to absorb and take up cadmium ions from solution, while the shoot/root ratios of QFS14 were higher than the other two cultivars (QFS14 > RH118 > No. 665).

Discussion

Cadmium pollution has become a major environmental problem with the development of modern industry and agriculture. Salt et al., (1995, 1996) proposed the use of metal-accumulating plants (e.g., Brassica juncea, Thlaspi caerulescens, Zea mays and Helianthus annuus) to remove toxic metals. According to the accepted shoot concentration defining hyperaccumulation being 0.01% (w/w) for Cadmium (Baker et al., 2000), the three cultivars of H. annuus can be considered metal hyperaccumulators (Table 1). Among three cultivars, Cd concentrations in the stem and leaves of RH118 were 149.2 μg/g DW (at $10^{-5}$ M Cd) and 963.1μg/g DW (at $10^{-4}$ M Cd) respectively, suggesting a greater ability to accumulate Cd when compared with QFS14. This cultivar, produced more roots and a higher biomass than No. 665, makes it a better choice for remediation of Cd contaminated soils.

The results in the present investigation provide basic information necessary for further development of phytoremediation methods and confirm previous studies (Lin et al., 2003; Madejón et al., 2003; Pena et al., 2006 and Soudek et al., 2006). Localization of Cd in plants depends on the total concentration of contaminant in soil or water. The results from this investigation showed that the Cd content in roots increased with increasing solution concentrations of Cd. Cadmium primarily accumulated in roots, and small amounts were transferred to shoots in the seedlings treated with $10^{-3}$ M Cd. The reported differences in root uptake and shoot accumulation might be explained by the fact that one of the normal functions of roots is to selectively remove ions from the soil solution (Salt et al., 1997). Cadmium-tolerant plants must be able to prevent the absorption of excess Cd or detoxify the Cd after it has been absorbed.

A few unfavourable changes in the appearance of the Cd-treated seedlings of H. annuus in the present investigation were observed, for example browning roots, the decreased number of roots, and growth inhibition of roots and shoots (Figs. 1 and 2). Chlorotic phenomena in seedlings treated with Cd in this investigation were also found, which agrees with the findings of Das et al., (1997).

It is likely that Cd tolerance mechanisms may differ depending on the species. More work still needs to be done to determine the mechanism of high uptake and accumulation of Cd by H. annuus. In our investigation, sunflower accumulated substantial amounts of Cd at $10^{-4}$ M Cd, but seedling growth was severely decreased. Therefore, it is important to select a suitable species with capabilities for high uptake and accumulation of Cd without the plant undergoing severe damage.

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References


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